

Appendix B

Hydrogeological Investigation

CITY OF HAMILTON

HYDROGEOLOGICAL ASSESSMENT – REV. 3 BLOCK 1, FRUITLAND-WINONA BLOCK SERVICING STRATEGY REPORT

MARCH 2024

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HYDROGEOLOGICAL ASSESSMENT – REV. 3 BLOCK 1, FRUITLAND-WINONA BLOCK SERVICING STRATEGY REPORT

CITY OF HAMILTON

PROJECT NO.: TP115082 MARCH 2024

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REVISION HISTORY

EXECUTIVE SUMMARY

WSP Canada Inc. (WSP) (formerly known as Wood Environment & Infrastructure Solutions) has carried out a hydrogeological assessment at Block 1 of the Fruitland-Winona Secondary Plan to fulfil the requirements of the Terms of Reference for the Fruitland-Winona Block Servicing Strategy (January 2014). WSP issued a draft Hydrogeological Assessment Report in November 2015. Subsequently, the hydrogeological assessment was updated in 2017 as part of a larger Servicing Strategies Report. Following comments, this Hydrogeological Assessment Report (Rev. 2) has been prepared to update the assessment based on the current understanding for the site.

The field work associated with the investigation consisted of installation of 6 monitoring wells, monitoring of groundwater levels, sampling and chemical analysis of groundwater from 3 monitoring wells, installation of pressure transducers in 4 monitoring wells, installation of pressure transducers in 4 monitoring wells, slug testing at 5 monitoring wells, and stream flow monitoring.

In the vicinity of Block 1, the surface topography is relatively flat with a gentle slope down towards the north, generally following the bedrock topography. The site is bordered by two permanent watercourses, Watercourse 5, which flows from south to north along the west edge of Block 1 (east of, but roughly parallel to Fruitland Road), and Watercourse 6, which flows from south to north along the east edge of Block 1 (east of Jones Road). While these are mapped as permanent watercourse features, observed flow in these features tended to be slow to intermittent.

Block 1 is located within the Iroquois Plain Physiographic Region (Chapman and Putnam, 1984) of Southern Ontario. According to Chapman and Putnam, 1984, the region of the Iroquois Plain to the west of Grimsby is characterized by heavy textured, low permeability soil developed on red clay derived from the underlying Queenston Formation. The Queenston Formation Shale is generally compact and dense with poor pore space interconnectivity and poor water yielding capabilities. During drilling on the site, bedrock was found to occur at depths from 1.0 to 2.2 m below ground surface. The surficial soil is identified as Halton Till, a clayey silt-clay till which is in agreement with the observations from boreholes drilled as part of the current field program.

Groundwater level monitoring using automatic pressure transducers indicated a trend of seasonal water level fluctuations with groundwater levels generally rising annually from February to April then generally decreasing between April and December. The range of water level fluctuations observed during the period from June 2015 to April 2017 was approximately 1.5 m at BH-2, 1.9 m at BH-1 and 3.6 m at BH-4.

Hydraulic conductivities determined from slug tests carried out in the monitoring wells ranged from 8.7x 10-5 m/s to 2.8 x 10-8 (geometric mean 8.5 x 10-7 m/s). The degree of variability in hydraulic conductivity is likely a reflection of the variability of the amount of weathering and fracturing occurring at different locations on the site. The groundwater flow direction is generally from the south-southwest towards the north-northeast with an average gradient of approximately 1.9%.

Groundwater sampling was carried out at monitoring wells BH/MW-1, BH/MW-2, and BH/MW-5 on August 4, 2015. In general, the water chemistry analyses show values typically found in groundwater derived from the Queenston Shale formation. The analysis results were compared with standards obtained from the Ontario Provincial Water Quality Objectives (PWQO) and from Table 7 – Non-potable groundwater, Generic Site Condition Standards for Shallow Soils in a Non-Potable Ground Water Condition. No values were obtained which exceed the Table 7 values. Results in excess of the PWQO were obtained for Boron and Uranium at all locations, and for Cobalt and Silver at BH/MW-1 and BH/MW-2. These results are likely naturally occurring as they are typical of the underlying Shale bedrock found in the area.

Surface water flow measurements and observations were carried out at Watercourse 5 and Watercourse 6 at two locations each (upstream – Regional Road 8 and downstream – Barton Street) on September 4, 2015. Low flows were measured at the upstream locations (12 m^3/day at Watercourse 5 and 6.9 m^3/day at Watercourse 6). Both watercourses were observed to be dry at the downstream locations.

Water balance calculations for Block 1 was carried out based on existing conditions and proposed postdevelopment conditions. Comparison between the pre- and post-development calculations indicates a reduction in evapotranspiration and infiltration and an increase in runoff volumes resulting from the increase in impervious surfaces. In order to address the deficit of infiltration due to development a number of LID measures can be used. The measures most likely to be implemented on this site would be downspout disconnection; increased topsoil depths (200 mm minimum); grassed swales to promote infiltration and TSS removal; infiltration trenches/swales (rear yard drainage swales with 150 mm topsoil rock gallery/storage median and perforated underdrain; soak away pits (rock filled galleries or chambers to store and infiltrate runoff; enhanced tree pits (enlarged chamber to receive direct runoff from streets); and bioswales (enhanced vegetative swale with filtration, attenuation and infiltration capabilities.

Development of Block 1 will tend to reduce the amount of infiltration of precipitation towards the water table primarily due to reduction in the amount of permeable area. Other factors which could contribute to this effect include increased compaction of the subsurface soils due to heavy vehicle traffic during construction, effects due to changes in site grading and changes in surface soils and vegetation type. Additionally, the excavation of trenches to accommodate underground utilities could create more permeable pathways for groundwater flow. Taken together, these factors would tend to result in a lowering of the water table in the vicinity of the development.

In the southern half of the Block 1 area, the low permeability of the surficial materials over which the watercourses flow (Halton Till) allow for very little interaction between the surface water and groundwater in these areas. Additionally, the watercourses do not transport large volumes of water and are observed to become dry during periods with low precipitation (such as during the summer months).

Groundwater levels near ground surface were measured at some locations during the field investigation, therefore it is likely that foundation drainage and sump pumps will be required for buildings having basements.

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TABLE OF **CONTENTS**

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TABLES

FIGURES

APPENDICES

1 INTRODUCTION

WSP Canada Inc. (WSP) carried out a hydrogeological assessment at Block 1 of the Fruitland-Winona Secondary Plan to fulfill the requirements of the Terms of Reference for the Fruitland-Winona Block Servicing Strategy (January 2014). WSP issued a draft Hydrogeological Assessment Report in November 2015. Subsequently, the hydrogeological assessment was updated in 2017 as part of a larger Servicing Strategies Report. Following comments, this Hydrogeological Assessment Report (Rev. 2) has been prepared to update the assessment based on the current understanding for the site.

Block 1 (the 'site') is located in the City of Hamilton and is bounded on the west by Fruitland Road, on the north by Barton Street on the south by Regional Road 8 and on the east by Watercourse 6, which runs south to north on the east side of Jones Road (Figure 1). Current land use is primarily residential and commercial or institutional along the roadways, with the interior of the block being open space (former agricultural land).

The field work associated with the investigation consisted of the following items:

- Installation of 6 monitoring wells;
- Monitoring of groundwater levels;
- Sampling of groundwater from 3 monitoring wells;
- Installation of pressure transducers in 4 monitoring wells;
- Slug testing at 5 monitoring wells; and
- Stream flow monitoring.

2 GEOLOGIC SETTING

In the vicinity of Block 1, the surface topography is relatively flat with a gentle slope down towards the north, generally following the bedrock topography. The site is bordered by two permanent watercourses, Watercourse 5, which flows from south to north along the west edge of Block 1 (east of, but roughly parallel to Fruitland Road), and Watercourse 6, which flows from south to north along the east edge of Block 1 (east of Jones Road). While these are mapped as permanent watercourse features, observed flow in these features tended to be slow to intermittent.

Block 1 is located within the Iroquois Plain Physiographic Region (Chapman and Putnam, 1984) of Southern Ontario, a lowland bordering Lake Ontario which was occupied by Lake Iroquois during the period when the last glaciation was receding but remained in the St. Lawrence Valley. According to Chapman and Putnam, 1984, the region of the Iroquois Plain to the west of Grimsby is characterized by heavy textured, low permeability soil developed on red clay derived from the underlying Queenston Formation.

The bedrock is identified as Queenston Formation Shale on Preliminary Map P.993, Quaternary Geology, Grimsby Area, Southern Ontario (Ontario Division of Mines, 1975. According to Singer et al. (2003), the Queenston Shale is generally compact and dense with poor pore space interconnectivity and poor water yielding capabilities. Data obtained from water well records indicate a relatively wide range of transmissivity values with 10 and 90 percentile values ranging between 0.5 and 27.9 m²/day respectively. In general, only the weathered zone in upper 3 to 5 m of the shale provide sufficient water supplies for domestic use.

According to Preliminary Map 2401, Bedrock Topography Series, Grimsby Area, Ontario Geological Survey, 1981, the bedrock slopes down towards Lake Ontario from approximately 92 m asl in the vicinity of Regional Road 8 to approximately 85 m asl in the vicinity of Barton Street. For comparison, the ground surface elevation varies from 93 to 88 m asl, indicating that the overburden is approximately 1 to 3 m thick.

On Quaternary Geology mapping prepared by the Ontario Geological Survey (2012), the area of Block 1 is identified as being an area of shallow bedrock, with some outcropping at or near ground surface, however the surficial soil of the surrounding area is identified as Halton Till: clayey silt-clay till (Figure 2), which is in agreement with the observations from boreholes drilled as part of the current field program. During drilling on the site, bedrock was found to occur at depths from 1.0 to 2.2 m below ground surface.

The Water Well Database of the Ontario Ministry of the Environment and Climate Change indicated at the time of the search (2015) that a total of 18 wells located within Block 1, of which 2 are recent 2" diameter plastic monitoring wells and the remainder being 6" diameter water wells drilled in the 1950's, primarily for domestic water supply (Figure 3).

The 2" diameter monitoring wells would likely have been installed as part of development, or other investigative work in the area. These wells are assumed to be abandoned if not still in use. The domestic water wells are primarily associated with residences located along the roads at the perimeter of the block and were generally drilled into the shale bedrock at depths of about 8 to 16 m below ground surface. According to the well records, the wells generally yielded fresh water at flow rates of about 9 to 23 L/min. It is unlikely that these wells are in current use for domestic water supply as the area is serviced by the municipal water supply system.

3 FIELD WORK

In conjunction with the geotechnical field work, a total of 6 monitoring wells, constructed of 50 mm diameter PVC screen and riser were installed from June 8 to June 10, 2015 in order to monitor water levels and obtain groundwater samples. The locations of the monitoring wells are shown in Figure 4. A seventh monitoring well, intended for the southern area adjacent to the cemetery along Regional Road 8, was not able to be installed due to site access limitations. Ground surface elevations at the borehole locations were determined from topographic mapping.

The monitoring wells consisted of 2 in. diameter, 1.5 m long screens placed at the bottom of each borehole to depths between 3.8 m and 5.2 m below ground surface and riser pipe extending to above ground surface. Each monitoring well was provided with a steel protective casing. Logs of the boreholes showing the position of the well screens are presented in Appendix A.

As indicated in the logs of boreholes, the site stratigraphy consists of a thin topsoil layer underlain by Silty Clay, which is in turn underlain by red weathered Shale. The shale was encountered at depths ranging from 1.07 m to 2.21 m below ground surface. In all cases the well screens were placed within the weathered shale unit.

The elevations of the wells were estimated using available topographic mapping, with ground surface elevations at each of the monitoring wells ranging between 87.4 m and 92.7 m.

3.1 WATER LEVELS

Water levels were obtained from all wells at the time of drilling on June 8, 9, and 10, 2015. Additional water levels were obtained on July 20 prior to well development, on July 21/22 prior to carrying out slug testing, and on August 4, and September 4 and 14, 2015. Pressure transducers were deployed in select monitoring wells in late 2015 to automatically record water levels in the monitoring wells over an extended period of time. The pressure transducers were installed in BH/MW-6 on August 5, 2015, in BH/MW 5 on August 31, 2015 and in BH/MW-1, BH/MW-2, BH/MW-4 on September 4, 2015. The transducers in BH/MW-5 and BH/MW-6 were removed on September 29, 2015 and the transducers in BH/MW-1, BH/MW-2, and BH/MW-4 remained in place until April 12, 2017.

Water level readings obtained in 2015 and 2017 are presented in **Table 1**.

Table 1: Water Levels Recorded in the Monitoring Wells

Note:

*Red text indicates water levels readings above ground surface

A hydrograph of water level elevations recorded by manual measurement and automatic pressure transducers programmed for hourly readings is presented as Figure 5. The hydrograph also shows daily precipitation as recorded by the Vineland weather station.

The locations of geological cross sections based on the units encountered during drilling are presented in Figure 7. The cross sections showing the site stratigraphy and groundwater level are presented in Figure 8.

In general, although short term periods of artesian conditions were noted in July 2015 in BH/MW-1, water levels were seen to be declining over the initial period of measurement during August to September 2015. The overall decline in water levels varied between the monitoring locations from about 0.5 m at BH/MW-2 to about 1.5 m at BH/MW-1 during this period. Figure 5 shows that water levels continued to decline at BH/MW-1 to about October 19, 2015, at BH/MW-2 to about December 26, 2015, and at BH/MW-4 to about January 2, 2016. At BH/MW-1 there was a slow rise in water level between October and the end of 2015 followed by a more rapid rise at all three locations where transducers were present until a high point was reached during the first half of April 2016.

The transducer records show a trend throughout 2016 and 2017 of rising water levels from January through to April, then lowering levels beginning in April. In 2016, the lowering trend at BH/MW-1 continues through to mid-December whereas BH/MW-2 is relatively stable from mid- July through to December. At BH/MW-4 the transducer appears to have been out of the water for a period between the end of October and mid-January and therefore the lowest level reached at this location was not recorded. The yearly magnitude of water level variation at BH/MW 1 is about 1.85 m during 2016, and at BH/MW 2 is about 1.34 m. BH/MW-4 showed the greatest magnitude of seasonal variation in water level with a value of greater than 3.5 m between the maximum and minimum levels. This amount of seasonal change at this location results in water levels at BH/MW-4 being higher than at BH/MW-2 during the period from February to July 2016 and lower than at BH/MW-2 during the remainder of the year. The same trend is evident beginning in February 2017 up to the point at which the transducers were removed in April.

The precipitation data collected at the Vineland Station RCS were compared to the 1961-1990, 1971-2000 and 1981-2010 Climate Normal Data available from Environment Canada's Vineland Rittenhouse Meteorological Station (ID: 6139143; location: 43°10'00.000" N, 79°25'00.000" W; elevation: 94.50 m), located approximately 30 km east of Block 1. This location was chosen for its similarity in elevation, topography and location with respect to Lake Ontario and the Niagara Escarpment when compared with other Meteorological Stations in the vicinity.

Comparison of the three "normal" annual precipitation values indicates a trend towards somewhat higher yearly precipitation totals as more recent data is incorporated. At the Vineland Station RCS, comparison of the 2015 and 2016 data with the Climate Normal data shows that overall these years were comparatively drier than the normal values. In addition, the totals for January to June for 2015 and 2016 were also drier than the normal totals for these months.

In contrast, the precipitation totals for the first 6 months of 2017 are significantly higher than the previous two years and also significantly higher than the Climate Normal data. In fact, the total precipitation recorded from January to June 2017 exceeds the total precipitation recorded in all of 2016.

It would be expected that peak water levels recorded at Block 1 during 2017 would be higher than those recorded during 2016, however the latest data recorded for 2017 appears to suggest that the peak values had been reached at about April 6-8 and had been dropping somewhat from then until the transducers were removed on April 12. It is possible, however that higher values were achieved at a later time and have not been recorded. The highest

recorded values during April 2017 are comparable to those observed during 2016 at BH/MW-1 and BH/MW-4 and were about 0.23 m higher at BH/MW-2.

3.2 SLUG TESTING

The monitoring wells were developed by pumping using Waterra tubing on July 20, 2015, followed by slug testing on July 21 and 22, 2015. After allowing the water levels in the monitoring wells to recover to their pre-pumping levels after development, rising head slug testing was carried out at each well location. Results of slug testing are presented in Table 2. Details of the slug test analyses are presented in Appendix B.

All the monitoring wells tested had screened intervals located within the shale bedrock, although BH/MW-1 had a short section within the overlying Clayey Silt/Silty Clay. As can be seen in Table 2, there was a wide range of hydraulic conductivity values determined from the shallow shale bedrock, with values varying over a range of more than three and a half orders of magnitude, from a low of 2.8×10^{-8} m/s at BH/MW-4 to 8.7×10^{-5} m/s at BH/MW-5. The average value obtained was 1.5 x 10⁻⁵ m/s and the geometric mean was 8.5 x 10⁻⁷ m/s. There does not appear to be any correlation between the magnitude of the hydraulic conductivity and the depth of the well screen below the bedrock surface or water level. The degree of variability in hydraulic conductivity is likely a reflection of the variability of the amount of weathering and fracturing occurring at different locations on the site.

3.3 GROUNDWATER FLOW

Approximate surface elevations for each of the borehole locations were used to calculate the approximate high groundwater elevation which was then plotted in order to estimate the direction of groundwater flow and gradient. Contours of groundwater elevation are presented in Figure 5. The figure shows that the groundwater direction is generally from the south-southwest towards the north-northeast with an average gradient of approximately 1.9 %.

The surface water features, such as the streams located along the west and east edges of the site, would serve as local groundwater discharge zones, with Lake Ontario to the north serving as the regional discharge zone. Additionally, any ponds located on the site may also function as either local groundwater discharge features, or alternatively, may collect local ponded surface water and allow it to infiltrate back into the groundwater system.

3.4 GROUNDWATER SAMPLING

Groundwater sampling was carried out at monitoring wells BH/MW-1, BH/MW-2, and BH/MW-5 on August 4, 2015. Prior to the collection of samples, a minimum of three well volumes of water were purged from each well in order to remove any stagnant water and ensure that water representative of that in the formation was obtained.

Prior to the collection of samples, field parameters (Temperature, pH, Conductivity, Total Dissolved Solids, and Salinity) were measured using a PCSTestr 35 water tester. The field parameters obtained are presented in Table 3. Samples were transported to an accredited laboratory and submitted for analysis of general chemistry parameters and metals. Samples submitted for analysis of metals were filtered in the field at the time of sampling. The results of the analyses are presented in Table 4. The Laboratory Certificate of Analysis is presented in Appendix C.

Table 3: Chemical Field Parameters

Sampling Date 04/08/2015 04/08/2015 04/08/2015 Published Standards Chemistry Parameters Depictive Chemistry Parameters Units BH/MW-1 BH/MW-2 BH/MW-5 PWQO Alternative Table 7 Sewer Bylaw Anion Sum **contract to the me/L** 33.6 64.1 30.5 Bicarb. Alkalinity (calc. as CaCO3) \vert mg/L 270 230 310 Calculated Parameters **Calculated Parameters** Calculated TDS mg/L 2,000 4,000 1,900 Carb. Alkalinity (calc. as CaCO3) \vert mg/L <1.0 <1.0 <1.0 1.0 Cation Sum me/L 28.3 61.7 32.1 Hardness (CaCO3) mg/L 1,200 2,000 1,000 Ion Balance (% Difference) % 8.65 1.89 2.61 Langelier Index (@ 20C) N/A 0.579 0.705 0.777 Langelier Index (@ 4C) \vert N/A \vert 0.335 0.464 0.534 $Saturation pH (@ 20C)$ N/A 6.76 6.70 6.78 Saturation pH (@ 4C) \vert N/A \vert 7.01 \vert 6.94 7.02 Total Ammonia-N mg/L 0.61 0.26 0.47 Unionized Ammonia (calculated) | ug/L | 3.28 | 1.75 | 4.24 | 20 ug/L² Conductivity umho/cm 2,900 5,400 2,800 Dissolved Organic Carbon 1 mg/L 2.4 2.9 2.6 Orthophosphate (P) mg/L <0.010 <0.010 <0.010 **Inorganics** pH pH 7.34 7.40 7.55 6.5 - 8.5 6.0 – 11.0 Dissolved Sulphate (SO₄) mg/L 1,100 2,000 860 1500³ Alkalinity (Total as CaCO3) mg/L 270 230 310 Dissolved Chloride (Cl) | mg/L | 160 | 670 | 230 | 1500³ | 1,800 Nitrite (N) | mg/L | <0.010 | <0.010 | <0.010 Nitrate (N) mg/L 1.27 <0.10 <0.10 Nitrate + Nitrite **mg/L** | mg/L | 1.27 | <0.10 | <0.10 Dissolved Aluminum (Al)
Dissolved Antimony (Sb) ug/L <0.50 <0.50 <0.50 <0.50 20 53,4 Dissolved Antimony (Sb) ug/L <0.50 <0.50 <0.50 <0.50 $\frac{1}{20}$ = $\frac{53.4}{16,000}$ Dissolved Arsenic (As) ug/L <1.0 <1.0 <1.0 5 13,4 1,500 Dissolved Barium (Ba) | ug/L | 69 | 39 | 49 | | | 23,000 Dissolved Beryllium (Be) ug/L <0.50 <0.50 <0.50 1,100 53 Dissolved Boron (B) ug/L 970 4,800 2,400 200 36,000 **Metals** Dissolved Cadmium (Cd) ug/L <0.10 <0.10 <0.10 1 13,4/0.0085 2.1 Dissolved Calcium (Ca) ug/L 280,000 490,000 230,000 Dissolved Chromium (Cr) ug/L <5.0 <5.0 <5.0 <5.0 53,4/0.085 640 Dissolved Cobalt (Co) ug/L 0.97 2.0 0.76/0.008 0.9 53,4 52 Dissolved Copper (Cu) ug/L 1.1 <2.0 <1.0 5 2^{3,4}/0.05⁵ 69 Dissolved Iron (Fe) \vert ug/L \vert <100 \vert <100 \vert 300 50^{3,4}

Table 4: Results of Chemical Analysis

Hydrogeological Assessment – Block 1, Fruitland Land-Winona Block Servicing Strategy Report WSP Canada Inc. Project No.: TP115082 March 2024
City of Hamilton Page 6 City of Hamilton Page 6

Notes:

PWQO Ontario Provincial Water Quality Objectives.

Table 7 Non-potable groundwater, Generic Site Condition Standards for Shallow Soils in a Non-Potable Ground Water Condition, Soil, Ground Water and Sediment Standards for Use Under Part XV.1 of the Environmental Protection Act, MOE, 2011.

¹ PWQO for phosphorus is a general guideline intended to limit excessive algal or plant growth in surface waters.

² PWQO for ammonia is based on the percentage of unionized ammonia which varies according to temperature and pH. Calculations based on these factors indicate that the percentage of unionized ammonia in these samples is below the PWQO.

- ³ Parameter measured in mg/L.
-

4 Analysis result refers to total metal.

⁵ Storm sewer limit, otherwise sanitary-combined limit.

(1) Detection Limit was raised due to matrix interferences.

In general, the water chemistry analyses show values typically found in groundwater derived from the Queenston Shale formation. According to Singer, et al (2003), the water quality varies from poor to good, with hardness generally hard to very hard and high values of Total Dissolved Solids, Sodium, Chloride, and Iron are commonly encountered.

In Table 4 previous versions of this reported were compared with standards obtained from the Ontario Provincial Water Quality Objectives (PWQO) and from Table 7 – Non-potable groundwater, Generic Site Condition Standards for Shallow Soils in a Non-Potable Ground Water Condition. In this report revision Table 4 also includes comparison of the results to the limits established in the City of Hamilton Sewer Bylaw. In order to establish the suitability of disposal of excess water during construction future groundwater sampling results will also be compared with the City of Hamilton Sewer Bylaw.

No values were obtained which exceed the Table 7 values. Results in excess of the PWQO were obtained for Boron and Uranium at all locations, and for Cobalt and Silver at BH/MW-1 and BH/MW-2. A result for Sulphate exceeded the limit for Sulphate at BH/MW-2. These results are likely naturally occurring as they are typical of the underlying Shale bedrock found in the area.

3.5 STREAM FLOW

Three surface water courses, designated Watercourse 5, Tributary to Watercourse 5, and Watercourse 6 occur within Block 1 (Figure 4).

Watercourse 5 drains northerly adjacent to Fruitland Road and discharges to Lake Ontario approximately 1.3 km downstream. The Tributary to Watercourse 5 drains to the north and discharges into a storm sewer on Barton Street. Watercourse 6 is located along the eastern boundary of Block 1 to the east of Jones Street. It also drains towards the north and discharges into Lake Ontario approximately 1.8 km to the north.

On September 4, 2015 the site was visited in order to attempt to measure stream flow volumes under low flow conditions. No precipitation had been recorded for a minimum period of at least 72 consecutive hours immediately preceding the collection of the stream flow measurements. The months of July and August had recorded unseasonably low amounts of precipitation compared to local climate normal data.

One upstream and one downstream location, adjacent to the north and south boundaries of Block 1 were investigated for each of Watercourse 5 and Watercourse 6. These locations are shown on Figure 4. Where water was observed at the measuring point station, a Hach 950 handheld flowmeter was used to estimate the flow rates in the watercourses.

The streambeds were observed to be dry at both of the Watercourse 5 and Watercourse 6 crossings at Barton Street on September 4, 2015.

At Watercourse 5, flow was observed and measured at the upstream end of Block 1, near Regional Road 8. At the location measured, water was observed to be flowing slowly in a channel 1.3 m wide and varying in depth from 1 to 10 cm. Based on measurements taken a flow rate of 1.4 x 10⁻⁴ m³/s (12 m³/day) was calculated.

At Watercourse 6, a low amount of flow was observed at the upstream side of Block 1 just north of the culvert at Regional Road 8. Flow was confined to a narrow channel between the rocky creek bed at this location and was calculated to be 8.0 x 10⁻⁵ m³/s (6.9 m³/day). Neither of these flows represent more than a trickle of water flowing onto the site.

3.6 GROUNDWATER RECHARGE AND DISCHARGE

In the vicinity of Block 1, the Niagara Escarpment serves as a major groundwater recharge feature. This is due to the change in elevation at the Escarpment face, as well as the karstic features in the bedrock that makes up the Escarpment, which allow for rapid vertical and horizontal movement of groundwater through the rock.

The proximity of the Escarpment to the south end of Block 1 can result in strong hydraulic gradients (rapid change in groundwater levels over a very short distance), which may result in higher groundwater levels and groundwater flows closer to the Escarpment, which typically lessen as the gradients become lower away from the Escarpment.

The watercourses that flow across the site showed little flow at the south end of the site (closer to the Escarpment) and were dry at the north end of the site. These watercourses originate on the Escarpment or from seeps along the Escarpment face and enter the site travelling over relatively low permeable material (Halton Till). As the watercourses move across the site, they move across bedrock units that are at or near ground surface (**Figure 3**). These bedrock units may be fractured and may be permeable enough to allow for some of the water to infiltrate. This, in combination with evaporation from the watercourse channel may result in the dry streams along the north end of the site.

Mapping of site surface water features suggest the presence of at least one pond on the site, located towards the southwest corner of the site and situated on the Halton Till. During the field investigation, no evidence of this pond was found. It is possible that it is a depression that seasonally fills with water and this water eventually evaporates or infiltrates, but additional monitoring would be required to determine the presence of this pond.

As the summer of 2015 was an unseasonably dry summer, this may have impacted both the observed flows in the streams, and the presence of any ponds or other potential wetland areas.

Generally, groundwater discharge zones would be expected to be found around areas of lower elevations and/or areas where rapid changes in elevation are found. While watercourses can also be indications of groundwater discharge zones, it is likely that the watercourses present at this site have little to no connection with the groundwater system as long as they are flowing across the low permeability Halton Till soils carrying water from the Escarpment area, however they may serve as groundwater recharge zones as they move across the bedrock units along the north half of the Block 1 area.

4 WATER BALANCE CALCULATION

A water balance calculation for Block 1 was carried out based on existing conditions and proposed postdevelopment conditions. Existing conditions were based on the classification of land use presented in Appendix D - Figure 3.2 - Ecological Land Classification, Appendix C – Natural Heritage Assessment Report, Fruitland Road Municipal Class Environmental Assessment (AECOM, 2011). Calculations for the proposed post-development conditions were based on areas of pervious and impervious conditions for the post- development land uses provided by Urbantech Consulting.

The existing water balance for Block 1 depends upon the current site conditions, such as location and climatic factors, soil texture, topography and type of vegetative cover and the extent of existing impervious cover, such as roads, driveways and buildings.

Potential impacts of development upon recharge to aquifers due to development of the property may occur due to reductions in pervious areas available for infiltration of precipitation, and changes to the soil texture, topography and vegetative cover. Installation of underground site services (sewers and watermains) may also act to intercept and redirect infiltrating water away from natural pathways.

4.1 METHODOLOGY

Recharge to aquifers occurs through a number of mechanisms, which may include direct infiltration from precipitation, infiltration from watercourses, agricultural irrigation, leakage from municipal utilities such as water mains and sewers, and flow between aquifers. Recharge is a complex process dependent upon the interaction of numerous factors such as rainfall duration and intensity, temperature fluctuations, wind velocity, duration of sunlight, soil texture and moisture holding capacity, depth of water table, topography and type of vegetation, all of which vary in both space and time. Precise measurement of the components of recharge is difficult to achieve in practice and therefore approximations and simplifications are commonly employed to characterize the water balance of a particular property.

A water balance is generally evaluated using an equation having the following form:

 $P = \Delta S + ET + RO + I$

Where

 P = precipitation ΔS = change in soil moisture storage ET= evapotranspiration RO = surface runoff I = infiltration

Meteorological data in the form of monthly average temperature and precipitation data, available from Environment Canada - Canadian Climate Normals 1981-2010 for the Vineland Rittenhouse Meteorological Station (43°10'00.000" N, 79°25'00.000" W, elevation 94.50 m), located approximately 30 km east of Block 1, was used to prepare the water balance.

While soil moisture storage fluctuates on a short term basis between precipitation events and seasonally due to cycles of freezing and thawing, the net change in storage over the long term is assumed to be zero. Calculation of evapotranspiration was carried out using a computer program, EVAP, which uses an accounting procedure based on the methodology originally presented by Thornthwaite (1948). Inputs to the computer model are mean monthly temperature, monthly total precipitation, latitude and hemisphere of the location of interest and soil moisture holding capacity of the soil based upon the soil properties and vegetation. The difference between the mean annual precipitation and the mean annual evapotranspiration is the annual water surplus, which is available to either infiltrate the soil surface or to flow over the land surface. A portion of the infiltration may flow vertically and percolate down to recharge groundwater, while a second component, termed interflow, may move laterally through shallow soils and re-emerge locally to the surface in low lying areas or stream beds. Both the interflow and direct runoff, which has not infiltrated the ground surface, together form the total surface runoff component.

Information regarding existing soil conditions at Block 1 was obtained from the observations made from samples collected during the field program. The logs of boreholes indicate that the surficial soils are brown to red and grey Silty Clay to Clayey Silt with a trace of Sand, Gravel and Shale at the borehole locations (Appendix A).

4.2 EXISTING CONDITIONS

Information regarding existing land use at Block 1 was obtained from examination of satellite imagery and from Appendix D - Figure 3.2 Ecological Land Classification, presented in Appendix C Natural Heritage, Municipal Class Environmental Assessment Study Phases 1 & 2 Report (AECOM, 2011 – Appendix D). The existing land use was classified based on a review of the above-referenced figure, available satellite imagery and on-site observations to determine approximate pervious and impervious areas for each land use type. Approximate areas for each land use type were calculated for the purpose of water balance calculations. The water balance for each of the existing land uses on Block 1 was calculated separately and then combined to produce the overall water balance for the site.

For the purpose of determining the soil moisture and infiltration factors for use in the calculations, the vegetative cover was categorized according to the descriptions provided in the above referenced Figure 3.2. Tables provided in the documentation accompanying the EVAP program were then used to determine the applicable soil moisture retention value to be used for each vegetative category. The vegetative categories and soil moisture retention values chosen for each category are presented in Table 5.

Table 5: Land Classification for Determination of Pervious Areas and Soil Moisture Values

In order to calculate the water balance for existing conditions, the separate land use categories were consolidated into four groups according to the assigned soil moisture storage values. The EVAP program was then run for each of the four consolidated groups to determine the actual evapotranspiration expected for each group. The total evapotranspiration for each group was calculated by multiplying the area of each group by the depth of actual evapotranspiration calculated by the EVAP program. In the case of the shallow rooted group, which includes both pervious (lawns) and impervious (buildings and parking lots) areas, the evapotranspiration was calculated by multiplying the EVAP value by the pervious area and assuming a 10% evaporation value for the impervious area.

A summary of the climatic water balance for existing conditions on Block 1 is provided i[n Table 6](#page-21-0). Total annual surplus is the amount of water available after evapotranspiration has occurred. Water surpluses of 315.1 mm/yr., 290.1 mm/yr. and 284.1 mm/yr. were calculated for shallow rooted, moderately rooted and deeply rooted vegetation, respectively and 273.1 mm/yr. for wooded areas.

Table 6: Existing Annual Water Balance for Existing Conditions on Block 1

Partitioning of water surplus between runoff and infiltration is dependent upon the soil type, texture, topography and type of vegetation cover. These are accounted for by the use of infiltration factors according to a method referenced in MOE (2003) and MOEE (1995) and are calculated by summing individual factors representing the contributions of topography, soil type and vegetative cover conditions. For Block 1, the factors are calculated for the individual land use types in order to calculate the individual volumes then summed to obtain the total volume for the area as a whole based on the proportional area of each land use. For the wooded areas, a topographic infiltration factor of 0.3, a soil infiltration factor of 0.2 and a vegetative cover factor of 0.2 was obtained from Table 2 in MOEE (1995). These factors were summed to obtain an infiltration factor of 0.7, meaning that 70% of the surplus will infiltrate. For the remaining areas the individual factors were 0.3, 0.2 and 0.1, for topography, soil, and vegetative cover factors respectively giving an infiltration factor of 0.6. Table 7 shows the resulting water balance for existing site conditions.

From Table 7 it is seen that under existing conditions, evapotranspiration is the largest component of the water balance, accounting for almost 57.5% of annual precipitation. Infiltration is 25.8 % of precipitation and runoff is 16.8%.

Table 7: Pre-Development Water Balance Volumes for Block 1

4.3 POST-DEVELOPMENT CONDITIONS

Development of property from rural and agricultural land use to urban residential and commercial/industrial land uses has effects on the natural water balance, primarily due to the addition of impervious surfaces, such as roads, parking lots, driveways and buildings. Impervious surfaces retain water at the surface, preventing infiltration into the soil, and removal of vegetation decreases the evapotranspiration component of the water balance within the affected area. Evaporation from impervious surfaces is a relatively minor component (estimated herein at 10% of precipitation) as compared to the 58% of precipitation occurring with existing vegetation. The net effect of the creation of impervious surfaces is therefore that most of the precipitation incident upon the area becomes surplus water and direct runoff, and natural infiltration is reduced.

Additional effects may also occur due to compaction of soil due to heavy vehicle traffic during construction, which acts to reduce the infiltration capacity of the underlying soils. Also, conversion of agricultural land or wooded areas to lawns affects the amount of evapotranspiration, as shallow rooted grasses draw water from less deep soil zones than trees and some agricultural crops. Runoff may also be reduced by reducing the surface gradient due to cut and fill operations during construction.

To calculate a post-development water balance for the Subject Property, areas of proposed post-development land uses and percent of the associated impervious areas for each type of land use were provided by Urbantech Consulting. The percentage of impervious area for each of the land use categories was then used in the calculation of infiltration and runoff for each proposed land use area. The soil moisture retention value was reduced to 100 mm for the pervious areas to account for the change in vegetation from the predevelopment vegetation types to grass (shallow rooted) when calculating evapotranspiration using the EVAP program. For the impervious areas, an evaporation factor of 10% was estimated to account for direct evaporation. The results of the post development calculations are presented in Table 8.

Table 8: Post-Development Water Balance for Block 1 (No Mitigation)

As can be seen in comparison with the results of the pre-development calculations presented in Table 7, the percentage of evapotranspiration has been reduced from 57.5% of precipitation to 30.9% due to the effects of development. Infiltration has also been reduced from 25.8% to 8.0% of precipitation and the largest component of the water balance is now surface runoff due to increasing the area of impervious surfaces.

The reduction of evapotranspiration and infiltration and resultant increase in runoff volumes would tend to reduce the amount of water available for local recharge to groundwater potentially lowering the local water table elevation and affecting the downstream groundwater gradients and flow directions. Surface water volumes, if allowed to flow towards surface watercourses would tend to increase due to the increase of runoff volumes. This could result in increased flows in watercourses downstream of the Site following development. This effect could be offset to some extent by decreasing contributions to surface flow from groundwater discharge to surface water bodies.

4.4 MITIGATION OF EFFECTS OF DEVELOPMENT ON THE WATER BALANCE

Potential impacts to the water balance from development may be minimized through the use of various low impact development (LID) measures for stormwater management to attempt to maintain the post-development infiltration volumes as close to the pre-development levels as possible. Mitigation measures may include such components as directing roof leaders to pervious areas, increasing topsoil thickness to retain more water in storage, soak away pits and dry wells, constructed infiltration galleries or trenches, pervious pipe systems, and grass swales.

Block 1 is underlain by primarily clayey silt soil and shale bedrock having relatively low hydraulic conductivity, which tends to limit the rate of infiltration. In response an LID storm water management system promoting the utilization of solutions that absorb, attenuate and exfiltrate storm water runoff rather than those that rely on infiltration would be appropriate. Notwithstanding the relatively low hydraulic conductivity of the soils, biofilters, rain gardens, permeable pavement and detention areas will achieve some degree of incidental infiltration. Despite the relatively low infiltration rates, the provision of storage within the LID management practices can attenuate a portion of the rainfall volume.

Despite the relatively low infiltration capacity of the soil, there is still potential for mitigation of the some of the effects of the increased impermeable area due to development and measures such as directing runoff from roofs, driveways and parking lots to bioretention areas, utilizing permeable pavement, implementing rainwater harvesting systems and increasing topsoil depth will be effective in attenuating runoff and improving water quality.

Infiltration targets for the proposed development have been provided by Urbantech based on the soil type and proposed land use (2.5 mm infiltration for sandy loam and all land uses, 1 mm for silty clay loam over residential, 2.5 mm for silty clay loam over commercial and institutional land use. For the total area of the development these targets result in a total infiltration requirement of 178,221 m³/yr. This represents an increase of the overall infiltration volume of approximately 104,000 m³/yr. The required infiltration targets for the site areas by soil type and land use type are presented in Table 9.

In order to address the deficit of infiltration due to development a number of LID measures can be used. The measures most likely to be implemented on this site would be downspout disconnection; increased topsoil depths

⁶¹⁷ Residential 1.19 Sandy Loam 10,610 2.5 ³⁰ 3,183

621 - Street B Residential | 0.19 | Sandy Loam | 1,694 | 2.5 | 30 | 508 Total 178,221

Residential | 1.09 | Silty Clay | 9,718 | 1.0 | 15 | 1,458

Ally Clay **19,080** 2.5 30 5,724

615 Institutional 2.14 Sandy Loam/

(200 mm minimum); grassed swales to promote infiltration and TSS removal; infiltration trenches/swales (rear yard drainage swales with 150 mm topsoil rock gallery/storage median and perforated underdrain; soak away pits (rock filled galleries or chambers to store and infiltrate runoff; enhanced tree pits (enlarged chamber to receive direct runoff from streets); and bioswales (enhanced vegetative swale with filtration, attenuation and infiltration capabilities.

5 POTENTIAL EFFECTS ON DEVELOPMENT

5.1 EFFECTS ON AQUIFERS AND WATER LEVELS

The subsurface materials at Block 1 consist of generally low permeability shale, a poor aquifer within its upper weathered zone, which is overlain by a low permeability till. There are no significant aquifers or recharge/ discharge zones present at Block 1.

Development of Block 1 will tend to reduce the amount of infiltration of precipitation towards the water table primarily due to reduction in the amount of permeable area. Other factors which could contribute to this effect include increased compaction of the subsurface soils due to heavy vehicle traffic during construction, effects due to changes in site grading and changes in surface soils and vegetation type. Additionally, the excavation of trenches to accommodate underground utilities could create more permeable pathways for groundwater flow. Taken together, these factors would tend to result in a lowering of the water table in the vicinity of the development.

5.2 EFFECTS ON SURFACE WATER

In the southern half of the Block 1 area, the low permeability of the surficial materials over which the watercourses flow (Halton Till) allow for very little interaction between the surface water and groundwater in these areas. The watercourses therefore primarily transport runoff from upstream areas (above the Escarpment and/or seeps at the Escarpment face). As the watercourses move across the site and encounter the bedrock units at or near ground surface, the surface water may infiltrate and the watercourses become dry. Additionally, the watercourses do not transport large volumes of water and are observed to become dry during periods with low precipitation (such as during the summer months).

Surface water volumes, if allowed to flow towards surface watercourses would tend to increase surface flow downstream of the Site, due to the increase of runoff volumes. This could result in increased flows in these watercourses following development. This effect could be offset to some extent by decreasing contributions to surface flow from any groundwater discharge to surface water bodies, however due to the factors outlined above, the groundwater contribution to surface flow is expected to be small.

6 FOUNDATION DRAIN FLOW

Groundwater levels were found near ground surface at some locations during the field investigation, therefore, it is likely that foundation drainage and sump pumps will be required for buildings having basements. The amount of flow to the drainage system will likely vary at different times and locations due to variations in the subsurface permeability across the area of the site and seasonal variations in water level. Flows are also expected to increase following large precipitation events and during the spring melt.

The amount of expected flow to the drainage system will depend on the length of the perimeter to be drained, the groundwater level, hydraulic conductivity of the subsurface materials and the amount of infiltrating precipitation intercepted by the drain.

Based on a building area of 2000 ft² (185.9 m²) and using a range of water levels and hydraulic conductivities observed at Block 1, groundwater flows to perimeter drainage are estimated to be in the range of 0.08 L/min to 4.3 L/min with a geometric mean of 0.9 L/min. These estimates are based on existing groundwater conditions and assume that groundwater storage is not depleted by the withdrawal of water. It is likely that the pore space and bedrock fractures of limited extent and that groundwater storage will become depleted as pumping occurs, thus limiting the amount of water to be removed. As groundwater storage is depleted, the water levels will decline and flows will be reduced from these values. Grading of lots to direct surface water away from building perimeters, perimeter drainage and underfloor drainage may help to mitigate the amount of seepage. There would also be a collective effect of multiple adjacent sump pumps operating simultaneously to draw down the water level in the vicinity of their operation. Due to the proximity of the Escarpment to the south of the property (a major source of groundwater recharge and high hydraulic gradients), it would be expected that houses located towards the south end of the site would be required to pump higher volumes than those located towards the north, where the gradients have lessened, resulting in lower groundwater flow rates.

It is generally recommended that sump drains should be designed to handle 30 L/min flow in low permeability conditions for each 1000 ft² of building area and 53 L/min for each 1000 ft² in permeable conditions (example at *abe-research.illinois.edu/pubs/factsheets/SumpPumps.pdf*).

7 CONSTRUCTION DEWATERING

Due to the presence of high water tables and variability of the hydraulic conductivity of the subsurface materials, some dewatering may be required during construction when excavating for basements and utility trenches. The amount of dewatering required will be dependent upon the size of the excavation, variations in hydraulic conductivity with location across the property and variations in water levels depending upon location and seasonal factors.

As an example, based on the assumptions made for inflows to sumps, flows to a typical residential excavation would experience the same range of flows of 0.08 o 4.3 L/min (115.2 to 6192 L/day) with a geometric mean of 0.9 L/min (1296 L/day).

Flows to trenches for construction of underground services are also expected to require some dewatering. Based on existing conditions, the geometric mean value for hydraulic conductivity and a 10 m long trench of 3 m width excavated to 2 m below the water table, groundwater inflow of about 4.17 m3/day (4,175 L/day) would be expected. Flows of this magnitude are expected to be able to be removed by sumps. Initial inflows could be higher, depending upon the hydraulic conductivity of the fractured shale in the vicinity of the excavation, however it is expected that flows will decline as water is removed from storage in the fractures. The extent of open trenches should be kept to the minimum required for the work to proceed in order to minimize the need to remove water from the excavation. If possible, scheduling of excavation works during periods of low groundwater level, such as late summer, would minimize dewatering requirements.

Placement of granular bedding materials may result in the creation of preferential flow paths for infiltrating surface water or groundwater. Anti-seepage collars should be provided at strategic points along sewers and other utility trenches, including between utility trenches and basements in order to prevent continuous flow along the backfilled excavations towards down gradient areas and basements and to retain the very low groundwater flows present under pre-development conditions. Collars should have an upper elevation or relief pipe to allow water to overtop the collars in the event of backup of water levels.

8 CONTINGENCY PLANNING

No significant aquifers were encountered during the field investigation and are not expected to be encountered during construction and development of the property. Groundwater is likely to be encountered during excavation for basements and utilities, but the amount of flow is expected to be low to moderate and to be able to be removed by pumping from sumps. Continuous pumping from sump pumps from residential basements is not expected to occur, however is expected to be more frequent in areas near surface water and at times of high water levels. No on site water supplies sewage disposal systems or surface and groundwater related infrastructure are known to be present.

Effects of development are expected to be a lowering of infiltration due to an increase in the area of impermeable ground cover such as roads, driveways and buildings, changes to grade, vegetative and soil cover and potential compaction of surface soil.

It is recommended that groundwater levels be monitored during the pre-construction and construction periods in order to be able to further assess the amount of natural seasonal fluctuation and the effect of construction on the groundwater levels at the property. During construction, it is recommended that any dewatering required for construction of basements or utility trenches be measured in order to assess the effect of dewatering.

9 CLOSURE

We trust that the information provided in this report is sufficient for your present purposes. Should you have any questions or concerns please do not hesitate to contact the undersigned.

Respectfully submitted,

WSP Canada Inc.

Prepared by: Reviewed by,

lilden

Hydrogeologist **Fellow Hydrogeologist** Fellow Hydrogeologist

Michael Anderson, M.Sc., P.Geo. Simon Gautrey, M.Sc., MBA, P.Geo., F.G.C.

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Wood Environment & Infrastructure Solutions 3450 Harvester Rd, Suite 100., Burlington, Ontario, L7N 3W5

wood.

tel: 905-335-2353 www.woodplc.com

Environment & Infrastructure Report.

SOURCE: ESRI ArcGIS Online, World Imagery, 2013.

Waterbody (OHN, 2012)

C:\Users\tana.yun\OneDrive - Wood PLC\Desktop\6. Small Tasks\FruitlandWinona Servicing_MAPS\Fig3_GW flow.mxd

Meters

FIGURE: 3

PROJECT N°: TP115082

SCALE: NTS

DATE: Nov 2021

X:\CA\CAFRE300-FRD\projectg\$\PROJECTS\TP115082_Fruitland_Winona_Servicing_Strategy\CAD\TP115082_FRUITLAND.dwg - 2/7/2024 11:06 AM - McCoy, Dale

X:\CA\CAFRE300-FRD\projectg\$\PROJECTS\TP115082_Fruitland_Winona_Servicing_Strategy\CAD\TP115082_FRUITLAND.dwg - 2/7/2024 11:06 AM - McCoy, Dale

Appendix A

Borehole Logs

Drilling Location: **4785460N; 606334E** Project Location: **Block 1, Stoney Creek, Ontario Example 2018** Logged by: 1, Stoney

Reviewed by: Compiled by: Revision No.: **TR KG TR** Date Completed: 8

Appendix B

Slug Test Analysis

Appendix C

Certificates of Analysis

Your Project #: 5500MKTG.010150 Site#: BLOCK 1 Your C.O.C. #: 521643-01-01 Site Location: FRUITLAND - WINONA

Attention:Michael Anderson

AMEC Foster Wheeler Environment & Infrastructure Hamilton - Standing Offer 505 Woodward Ave Unit 1 Hamilton, ON L8H 6N6

> **Report Date: 2015/08/13** Report #: R3627003 Version: 1 - Final

CERTIFICATE OF ANALYSIS

MAXXAM JOB #: B5F4677 Received: 2015/08/05, 17:10

Sample Matrix: Water # Samples Received: 3

Reference Method suffix "m" indicates test methods incorporate validated modifications from specific reference methods to improve performance.

* RPDs calculated using raw data. The rounding of final results may result in the apparent difference.

(1) Dissolved Organic Carbon (DOC) present in the sample should be considered as non-purgeable DOC.

(2) Values for calculated parameters may not appear to add up due to rounding of raw data and significant figures.

Your Project #: 5500MKTG.010150 Site#: BLOCK 1 Your C.O.C. #: 521643-01-01 Site Location: FRUITLAND - WINONA

Attention:Michael Anderson

AMEC Foster Wheeler Environment & Infrastructure Hamilton - Standing Offer 505 Woodward Ave Unit 1 Hamilton, ON L8H 6N6

> **Report Date: 2015/08/13** Report #: R3627003 Version: 1 - Final

CERTIFICATE OF ANALYSIS

MAXXAM JOB #: B5F4677 Received: 2015/08/05, 17:10

Encryption Key

Ma

Please direct all questions regarding this Certificate of Analysis to your Project Manager. Marijane Cruz, Senior Project Manager Email: MCruz@maxxam.ca Phone# (905)817-5756 ==

Maxxam has procedures in place to guard against improper use of the electronic signature and have the required "signatories", as per section 5.10.2 of ISO/IEC 17025:2005(E), signing the reports. For Service Group specific validation please refer to the Validation Signature Page.

Total Cover Pages : 2

RCAP - COMPREHENSIVE (WATER)

Lab-Dup = Laboratory Initiated Duplicate

N/A = Not Applicable

RCAP - COMPREHENSIVE (WATER)

RDL = Reportable Detection Limit

QC Batch = Quality Control Batch

Lab-Dup = Laboratory Initiated Duplicate

(1) Detection Limit was raised due to matrix interferences.

RCAP - COMPREHENSIVE (WATER)

Collected: 2015/08/04

TEST SUMMARY

Maximizing Collected: 2015/08/04 **Shipped: Received:** 2015/08/05

Page 6 of 12

Maxxam Analytics International Corporation o/a Maxxam Analytics 6740 Campobello Road, Mississauga, Ontario, L5N 2L8 Tel: (905) 817-5700 Toll-Free: 800-563-6266 Fax: (905) 817-5777 www.maxxam.ca

Test Description

Anion and Cation Sum Total Ammonia-N

AMEC Foster Wheeler Environment & Infrastructure Client Project #: 5500MKTG.010150 Site Location: FRUITLAND - WINONA Sampler Initials: MA

Maxxam ID: ATC709 **Collected:** 2015/08/04

Received: 2015/08/05

Shipped:

TEST SUMMARY

Orthophosphate **KONE** 4140166 N/A 2015/08/09 Alina Dobreanu Sat. pH and Langelier Index (@ 20C) CALC 4134703 N/A 2015/08/13 Automated Statchk Sat. pH and Langelier Index (@ 4C) CALC 4134704 N/A 2015/08/13 Automated Statchk Sulphate by Automated Colourimetry $XONE$ 4140164 N/A 2015/08/10 Alina Dobreanu Total Dissolved Solids (TDS calc) CALC 4134702 N/A 2015/08/13 Automated Statchk

GENERAL COMMENTS

Each temperature is the average of up to three cooler temperatures taken at receipt

Package 1 3.0°C

Results relate only to the items tested.

Maxxam Job #: B5F4677 Maxxam Job #: B5F4677
Report Date: 2015/08/13 **COMPUTE REPORT**

AMEC Foster Wheeler Environment & Infrastructure Client Project #: 5500MKTG.010150

Sampler Initials: MA Site Location: FRUITLAND - WINONA

Page 9 of 12

Maxxam Job #: B5F4677

Maxxam Job #: B5F4677
Report Date: 2015/08/13
Report Date: 2015/08/13

AMEC Foster Wheeler Environment & Infrastructure Client Project #: 5500MKTG.010150

Sampler Initials: MA Site Location: FRUITLAND - WINONA

N/A = Not Applicable

Duplicate: Paired analysis of a separate portion of the same sample. Used to evaluate the variance in the measurement.

Matrix Spike: A sample to which a known amount of the analyte of interest has been added. Used to evaluate sample matrix interference.

Spiked Blank: A blank matrix sample to which a known amount of the analyte, usually from a second source, has been added. Used to evaluate method accuracy.

Method Blank: A blank matrix containing all reagents used in the analytical procedure. Used to identify laboratory contamination.

NC (Matrix Spike): The recovery in the matrix spike was not calculated. The relative difference between the concentration in the parent sample and the spiked amount was too small to permit a reliable recovery calculation (matrix spike concentration was less than 2x that of the native sample concentration).

NC (Duplicate RPD): The duplicate RPD was not calculated. The concentration in the sample and/or duplicate was too low to permit a reliable RPD calculation (one or both samples < 5x RDL).

(1) Recovery or RPD for this parameter is outside control limits. The overall quality control for this analysis meets acceptability criteria.

VALIDATION SIGNATURE PAGE

The analytical data and all QC contained in this report were reviewed and validated by the following individual(s).

Ewa Pranjic, M.Sc., C.Chem, Scientific Specialist

Maxxam has procedures in place to guard against improper use of the electronic signature and have the required "signatories", as per section 5.10.2 of ISO/IEC 17025:2005(E), signing the reports. For Service Group specific validation please refer to the Validation Signature Page.

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Appendix D

Figures by Others

Natural Heritage Assessment of Lands Bounded by Fruitland Road, Glover Road, Barton Street and Highway 8, City of Hamilton **Figure 3.2 Ecological Land Classification**

Appendix E

Limitations

Limitations

- 1. The work performed in the preparation of this report and the conclusions presented are subject to the following:
	- a. The Standard Terms and Conditions which form a part of our Professional Services Contract;
	- b. The Scope of Services;
	- c. Time and Budgetary limitations as described in our Contract; and
	- d. The Limitations stated herein.
- 2. No other warranties or representations, either expressed or implied, are made as to the professional services provided under the terms of our Contract, or the conclusions presented.
- 3. The conclusions presented in this report were based, in part, on visual observations of the Site and attendant structures. Our conclusions cannot and are not extended to include those portions of the Site or structures, which are not reasonably available, in WSP's opinion, for direct observation.
- 4. The environmental conditions at the Site were assessed, within the limitations set out above, having due regard for applicable environmental regulations as of the date of the inspection. A review of compliance by past owners or occupants of the Site with any applicable local, provincial or federal bylaws, orders-incouncil, legislative enactments and regulations was not performed.
- 5. The Site history research included obtaining information from third parties and employees or agents of the owner. No attempt has been made to verify the accuracy of any information provided, unless specifically noted in our report.
- 6. Where testing was performed, it was carried out in accordance with the terms of our contract providing for testing. Other substances, or different quantities of substances testing for, may be present on-site and may be revealed by different or other testing not provided for in our contract.
- 7. Because of the limitations referred to above, different environmental conditions from those stated in our report may exist. Should such different conditions be encountered, WSP must be notified in order that it may determine if modifications to the conclusions in the report are necessary.
- 8. The utilization of WSP's services during the implementation of any remedial measures will allow WSP to observe compliance with the conclusions and recommendations contained in the report. WSP's involvement will also allow for changes to be made as necessary to suit field conditions as they are encountered.
- 9. This report is for the sole use of the party to whom it is addressed unless expressly stated otherwise in the report or contract. Any use which any third party makes of the report, in whole or the part, or any reliance thereon or decisions made based on any information or conclusions in the report is the sole responsibility of such third party. WSP accepts no responsibility whatsoever for damages or loss of any nature or kind suffered by any such third party as a result of actions taken or not taken or decisions made in reliance on the report or anything set out therein.
- 10. This report is not to be given over to any third party for any purpose whatsoever without the written permission of WSP.
- 11. Provided that the report is still reliable, and less than 12 months old, WSP will issue a third-party reliance letter to parties that the client identifies in writing, upon payment of the then current fee for such letters. All third parties relying on WSP's report, by such reliance agree to be bound by our proposal and WSP's standard reliance letter. WSP's standard reliance letter indicates that in no event shall WSP be liable for
any damages, howsoever arising, relating to third-party reliance on WSP's report. No reliance by any party is permitted without such agreement.